

ردیابی آلودگی آبهای زیرزمینی در محل دفن زباله ها با استفاده از فناوری ایزوتوپیهای پایدار ($\delta^{13}\text{C}$ و $\delta^2\text{H}$, $\delta^{18}\text{O}$)

حسین محمدزاده^{۱،*}، یان کلارک^۲

۱- مرکز تحقیقات آبهای زیرزمینی (متاب)، دانشگاه فردوسی مشهد، mohammadzadeh@alummi.uottawa.ca

۲- گروه علوم زمین، دانشکده علوم، دانشگاه اتاوا - کانادا

چکیده

در این مقاله، تاثیر شیرابه زباله ها بر روی سفره های آبهای زیرزمینی شهر اتاوا - پایتخت کانادا - با استفاده از ایزوتوپیهای پایدار ($\delta^{13}\text{C}$ و $\delta^2\text{H}$, $\delta^{18}\text{O}$) مورد ارزیابی قرار گرفته و این فناوری، آلودگی آبهای زیرزمینی توسط شیرابه زباله ها را تایید مینماید. مقادیر عددی بالای $^{13}\text{C}_{\text{DIC}}$ آبهای زیرزمینی (به ترتیب -6.5% و -1.0% برای سفره های کم عمق و عمیق) در مقایسه با $^{13}\text{C}_{\text{DIC}}$ آبهای زیرزمینی در بالادست محل دفن زباله ها (-15.2%) تاییدی بر تاثیر شیرابه زباله ها ($+8.8\%$ تا $+10.7\%$) بر روی آبهای زیرزمینی منطقه می باشد. عدم انطباق آبهای زیرزمینی بر روی خطوط اختلاط در دیگرامهای ترکیبی DIC با $\delta^{13}\text{C}_{\text{DIC}}$ و DOC با $\delta^{13}\text{C}_{\text{DOC}}$ نشان دهنده واکنشهای بیوشیمیایی مرتبط با کربن ارگانیکی محلول (DOC) می باشد که به نوبه خود بر روی غلظت DOC، غلظت کربن غیرارگانیکی محلول (DIC) و ایزوتوپ کربن ۱۳ تاثیر می گذارد. بنابراین علاوه بر اختلاط شیرابه زباله ها با آبهای زیرزمینی، واکنشهای بیوشیمیایی نیز بر کیفیت آبهای زیرزمینی تاثیر گذاشته است.

واژگان کلیدی: شیرابه زباله ها، ایزوتوپ پایدار ($\delta^{13}\text{C}$ و $\delta^2\text{H}$, $\delta^{18}\text{O}$)، کربن ارگانیکی محلول (DOC)، کربن غیرارگانیکی محلول (DIC)

Tracing Groundwater Contamination at Landfill Sites Using Stable Isotope Technique ($\delta^{18}\text{O}$, $\delta^2\text{H}$ and $\delta^{13}\text{C}$)

Abstract

In this paper, the stable isotope ($\delta^{18}\text{O}$, $\delta^2\text{H}$ and $\delta^{13}\text{C}$) was used to recognize groundwater contamination at Ottawa landfill site. The enriched $^{13}\text{C}_{\text{DIC}}$ (averages of 6.4% , -1.0% for the shallow, deep aquifers, respectively) in comparison to the upgradient pristine groundwater (-15.2%) confirm the leachate ($+8.8\%$ and $+10.7\%$) impact on these aquifers. Biochemical reaction of DOC in groundwaters is confirmed by deviation of groundwater samples from the mixing lines on DIC vs. $\delta^{13}\text{C}_{\text{DIC}}$ and DOC vs. $\delta^{13}\text{C}_{\text{DOC}}$ diagrams. Therefore, mixing and the reaction are two important processes which effected on groundwater quality at landfill site.

Keywords: Landfill leachate, stable isotopes ($\delta^{18}\text{O}$, $\delta^2\text{H}$ and $\delta^{13}\text{C}$), dissolved organic carbon (DOC), dissolved inorganic carbon (DIC)



Introduction

Leachate contains a mixture of dissolved organic constituents which can decay and dissolve through bacteria-mediated reactions resulting in generating different major carbon pools, such as DIC, DOC, and CH₄. The biogeochemical processes and outgassing of CO₂ can produce characteristic ¹³C signatures in the major carbon pools and also generate ²H enriched of water in landfill leachate (Baedecker and Back, 1979; Hackley et al., 1996; Christensen et al., 2001; Van Breukelen et al., 2004; North et al., 2004; Mohammadzadeh and Clark, 2008). The objective of this paper is to investigate groundwater contamination at Trail Road Landfill (TRL) site, located approximately 25 km west of the Ottawa City, using stable isotope signatures.

Site characterization and analytical methods

There are both a shallow and a deep aquifer at TRL area. Several individual and nested piezometers representing more than 200 monitoring wells were installed by City of Ottawa to sample the groundwater. The nested piezometers were screened at different depths including the shallow aquifer, the upper deep aquifer, and the lower deep aquifer. (Dillon Consulting Ltd., 2006). Leachate samples were taken from M32 monitoring well and groundwater samples were taken from multilevel monitoring wells located down-gradient and up-gradient of the TRL site. The detailed description of sampling procedures and analytical techniques is given by Mohammadzadeh et al. (2005) and Mohammadzadeh and Clark. (2008).

Discussion and conclusions

All results are tabulated in Table 1. Review of the available geochemical data from M57 and M120, as reference groundwater, shows no evidence of leachate influence at these locations.

The presence alone of methane at some locations (M4-1, M4-2, M37-2, M23-1, and M23-2) is evidence of leachate impact on groundwater. Further evidence is provided by ¹³C of DIC and that of CH₄. In comparison with pristine groundwater, samples from the shallow aquifer show variable concentrations of methane. Associated $\delta^{13}\text{C}_{\text{CH}_4}$ values range from -60 ‰, indicating biogenic production, to -32 ‰ indicating subsequent oxidation. DIC also carries evidence for leachate impact, with high and variable concentrations as compared to pristine groundwater, with $\delta^{13}\text{C}_{\text{DIC}}$ values that show variable additions of ¹³C-depleted DIC generated through the oxidation of methane. The deep aquifer monitored further from the landfill (M23 and M16), also has been affected by the landfill leachate. Similarly, the $\delta^{13}\text{C}_{\text{DIC}}$ in the upper (M37-2 and M23-2) and lower (M23-1 and M16-1) part of the deep aquifer shows variable degrees of both leachate impact on groundwater and subsequent methane oxidation. The $\delta^{13}\text{C}_{\text{DIC}}$ values at the lower part of the deep aquifer (M23-1 and M16-1) are enriched, 11.8 ‰ and -8.0 ‰, respectively, and the enriched values of $\delta^{13}\text{C}_{\text{CH}_4}$ at M16-1 clearly provide evidence for methane oxidation at this point.

No methane and a low amount of DOC are present in up-gradient (M57) and down-gradient (M120) background groundwater. Concentrations of DOC and CH₄ in M32 (5130 mg l⁻¹ and 1.6 mg l⁻¹, respectively) signify the input of the leachate plume from the unlined parts of landfill, and decrease in the groundwater flow direction. The high concentration of these species in M77 is due to the impacts of the adjacent Nepean landfill leachate on this part of the aquifer. Also decreasing along the flow direction is the ¹³C value of DIC, from the enriched value for the landfill site at M32 through M16-1, M114, M77-1, M77-2, M93 and M64. This reflects both dilution with background DIC ($\delta^{13}\text{C}$ of -17 ‰) plus contributions from the oxidation of dissolved organics and methane.



None of the average data obtained from samples taken from these aquifers plot on the mixing lines generated for the DIC and DOC vs. $\delta^{13}\text{C}$ diagrams (Figure not shown here), indicating reactive evolution in addition to simple dilution. Moreover, while DIC shows largely an increase in $\delta^{13}\text{C}$, DOC experiences a marked depletion in ^{13}C with evolution.

The groundwaters samples from both the shallow and upper deep aquifers plot on or below the OMWL on $\delta^2\text{H} - \delta^{18}\text{O}$ diagram (Figure is not shown here). The absence of a deuterium excess in these waters reflects both dilution and lack of significant methane production. Shifts in $\delta^2\text{H}$ due to methanogenesis is strictly occurs in unsaturated zones with the landfill where high organic carbon loads coupled with low water to gas and solids ratios imparts a significant enrichment of ^2H on the residual water.

At Trail Road Landfill (TRL) site, both the shallow and deep aquifers have been impacted by leachate infiltrating from the unlined portions of the landfill site. The concentrations and $\delta^{13}\text{C}$ of dissolved methane, the concentration and $\delta^{13}\text{C}$ of DIC, provide evidence for landfill leachate impact. Bulk DOC concentrations and $\delta^{13}\text{C}_{\text{DOC}}$ in groundwaters are affected by a combination of dilution and continued reaction. Consumption of DOC via acetate fermentation and oxidation of DOC within the aquifers is documented by: 1) deviation of groundwater samples from mixing lines, drawn for DIC vs. $\delta^{13}\text{C}_{\text{DIC}}$ and DOC vs. $\delta^{13}\text{C}_{\text{DOC}}$; 2) correlation between net reacted (lost) DOC and the net gains in for shallow and deep aquifers.

References

- Baedecker, M.J., Back, W., 1979. Hydrogeological processes and chemical reactions at a landfill. *Groundwater*, Vol. 17, pp. 429-437.
- Christensen, T.H., Kjeldsen, P., Bjerg, P.L., Jensen, D.L., Christensen, J.B., Baun, A., Albrechtsen, H.J. Heron, G., 2001. Review biogeochemistry of landfill leachate plumes. *Applied Geochemistry* Vol. 16, pp. 659-718.
- Dillon Consulting Limited, 2006. Trail Road and Nepean landfill sites monitoring and operating reports, report to Regional Municipality of Ottawa – Carleton.
- Hackley, K.C., Liu, C.L., Coleman, D.D., 1996. Environmental isotope characteristics of landfill leachate and gases. *Groundwater*, Vol. 34, No. 5, pp. 827-836.
- Mohammadzadeh, H., Clark, I.D., Marschner, M., St-Jean, G., 2005. Compound specific isotopic analysis (CSIA) of landfill leachate DOC Components. *Chemical Geology*, Vol. 218, pp. 3-13.
- Mohammadzadeh, H., Clark, I.D., 2008. Degradation pathways of dissolved carbon in landfill leachate traced with compound-specific ^{13}C isotope analysis of DOC. *Journal of Isotopes in Environmental and Health Studies*, Vol. 44 (3), pp. 267-294.
- North, J. C., Frew, R. D., and Peake, B. M., 2004. The use of carbon and nitrogen isotope ratios to identify landfill leachate contamination: Green Island Landfill, Dunedin, New Zealand. *Environmental International*, Vol. 30, pp. 631-637.
- Van Breukelen, B.M., J., Griffioen, J., Roling, W.F.M., Van Verseveld, H.W., 2004. Biogeochemistry and isotope geochemistry of a landfill leachate plume. *Journal of Contaminant Hydrology*, vol. 70, pp. 249- 269.



Table 1: Geochemical and isotopic data for the groundwater samples collected at TRL site. All concentrations in mg l⁻¹ and isotopic values in ‰ VPDB (for ¹³C) and in ‰ VSMOW (for ¹⁸O and ²H), if not stated.

| Sample ID | Date | H ₂ O | | DIC | | DOC | | CH ₄ | | CO ₂ | | α ¹³ C _{CO2-CH4} |
|------------------------|---------------|-------------------|------------------|--------------------|----------------------------------|--------------------|----------------------------------|--------------------|----------------------------------|----------------------------------|----------------------------------|--------------------------------------|
| | | δ ¹⁸ O | δ ² H | mg l ⁻¹ | δ ¹³ C _{DIC} | mg l ⁻¹ | δ ¹³ C _{DOC} | mg l ⁻¹ | δ ¹³ C _{CH4} | δ ¹³ C _{CO2} | δ ¹⁸ O _{CO2} | |
| M57 | July 22, 2005 | -13.6 | -96.9 | 48 | -13.4 | 0.9 | -26.8 | 0.00 | -50.8 | -17.9 | -0.4 | 1.035 |
| M57 | July 9, 2004 | -13.4 | -96.3 | 58 | -17.1 | 1.21 | -31.1 | 0.00 | bl | bl | | |
| Average | | -13.5 | -96.6 | 53 | -15.2 | 0.9 | -28.9 | 0.00 | -50.8 | -17.9 | -0.4 | 1.0 |
| Reference | | | | | | | | | | | | |
| M120 | July 21, 2005 | -9.8 | -78.2 | 59 | -13.3 | 0.6 | -26.9 | 0.00 | -47.4 | -17.2 | 11.5 | 1.032 |
| M120 | July 9, 2004 | -9.6 | -75.4 | 45 | -14.8 | 1.32 | -41.1 | 0.00 | bl | bl | | |
| M120 | Nov 14, 2003 | -9.4 | -74.4 | 47 | -14.3 | 0.5 | -22.1 | 0.00 | | | | |
| Average | | -9.6 | -76.0 | 50 | -14.1 | 0.5 | -30.0 | 0.00 | -47.4 | -17.2 | 11.5 | 1.0 |
| Pond | | | | | | | | | | | | |
| M79 | Nov 14, 2003 | -11.3 | -80.1 | 104 | -16.5 | 8.7 | -27.3 | 0.01 | | | | |
| DWP-SE | July 21, 2005 | -10.8 | -77.9 | 72 | -4.3 | 1.4 | | 0.21 | -9.9 | -12.0 | -8.6 | 0.998 |
| DWP-SE | July 9, 2004 | -10.9 | -66.1 | 85 | -7.3 | 2.5 | -41.6 | 0.27 | -44.5 | -13.7 | | 1.032 |
| DWP-SE | Nov 14, 2003 | -11.0 | -82.1 | 102 | -6.9 | 1.8 | -21.1 | 0.07 | | | | |
| DWP-NW | Nov 14, 2003 | -10.6 | -81.0 | 89 | -8.2 | 1.6 | -23.9 | 0.02 | | | | |
| Average | | -10.9 | -76.8 | 87 | -6.7 | 1.8 | -28.9 | 0.14 | -27.2 | -12.8 | -8.6 | 1.0 |
| Shallow Aquifer | | | | | | | | | | | | |
| M16-3 | July 8, 2004 | -12.8 | -87.5 | 68 | -17.5 | 2.0 | -32.4 | 0.00 | -74.8 | bl | | |
| M37-3 | July 20, 2005 | -11.6 | -77.4 | 247 | 6.1 | 30.8 | -26.9 | 0.26 | -32.5 | -4.2 | 8.0 | 1.029 |
| M37-3 | July 8, 2004 | -11.5 | -78.0 | 417 | 3.9 | 76.8 | -25.2 | 0.63 | -43.5 | -2.1 | | 1.043 |
| M37-3 | Nov 6, 2003 | -11.3 | -80.2 | 178 | -8.8 | 39.2 | -26.1 | 0.02 | | | | |
| M23-3 | July 8, 2004 | -12.6 | -86.9 | 49 | -20.8 | 2.6 | -28.8 | 0.00 | -61.8 | bl | | |
| M43 | July 8, 2004 | -12.2 | -85.9 | 332 | 18.8 | 18.5 | -18.8 | 6.28 | -57.9 | 18.0 | | 1.081 |
| M4-2 | July 22, 2005 | -11.8 | -85.3 | 151 | -11.9 | 3.9 | -26.5 | 7.27 | -69.0 | -15.2 | 9.2 | 1.058 |
| M4-2 | July 7, 2004 | -12.0 | -83.5 | 194 | -14.8 | 5.5 | -30.1 | 9.99 | -59.0 | -15.4 | | 1.046 |
| M4-2 | Nov 6, 2003 | -11.9 | -86.6 | 165 | -6.1 | 6.1 | -29.9 | 4.53 | | | | |
| M33-1 | July 7, 2004 | -11.3 | -79.3 | 339 | 1.6 | 20.2 | -25.6 | 3.04 | -49.6 | -0.8 | | 1.051 |
| M116-3 | Nov 7, 2003 | -13.2 | -98.3 | 47 | -13.6 | 2.6 | -34.1 | 0.00 | | | | |
| M86 | Nov 7, 2003 | -11.7 | -82.4 | 53 | -14.0 | 1.8 | -31.4 | 0.00 | | | | |
| Average | | -12.0 | -84.3 | 187 | -6.4 | 17.5 | -28.0 | 2.67 | -56.0 | -3.3 | 8.6 | 1.051 |

bl = blow detection limit

Table 1: Continued.

| Sample ID | Date | H ₂ O | | DIC | | DOC | | CH ₄ | | CO ₂ | | $\alpha^{13}C_{CH_4-CO_2}$ |
|----------------|---------------|------------------|--------------|--------------------|----------------------|--------------------|----------------------|--------------------|-----------------------|-----------------------|-----------------------|----------------------------|
| | | $\delta^{18}O$ | δ^2H | mg l ⁻¹ | $\delta^{13}C_{DIC}$ | mg l ⁻¹ | $\delta^{13}C_{DOC}$ | mg l ⁻¹ | $\delta^{13}C_{CH_4}$ | $\delta^{13}C_{CO_2}$ | $\delta^{18}O_{CO_2}$ | |
| M114-2 | July 21, 2005 | -11.7 | -84.0 | 38 | -13.6 | 1.3 | -26.7 | 0.00 | -59.6 | -17.9 | 7.5 | 1.044 |
| M114-2 | July 9, 2004 | | | 49 | -19.4 | 1.7 | -30.3 | 2.16 | -50.4 | | | |
| M107-2 | July 22, 2005 | -10.8 | -79.2 | 83 | -16.8 | 3.7 | -27.9 | 0.00 | -57.1 | -21.3 | -8.3 | 1.038 |
| M34 | July 22, 2005 | -11.7 | -85.2 | 214 | 4.7 | 3.6 | -38.1 | 0.68 | -39.7 | -2.2 | 4.9 | 1.039 |
| M64 | July 21, 2005 | -11.7 | -83.7 | 111 | -1.5 | 3.4 | -29.0 | 1.85 | -41.8 | -8.0 | -8.0 | 1.035 |
| M64 | July 9, 2004 | -11.8 | -83.0 | 143 | -7.6 | 4.8 | -35.3 | 3.61 | -43.6 | -11.1 | | 1.034 |
| M64 | Nov 14, 2003 | -11.7 | -82.0 | 151 | -1.7 | 3.2 | -33.9 | 1.21 | | | | |
| M77-2 | July 21, 2005 | -11.2 | -80.9 | 129 | -3.6 | 1.7 | -27.3 | 0.25 | -32.5 | -9.3 | 10.0 | 1.024 |
| M77-2 | July 9, 2004 | -11.4 | -79.6 | 166 | -9.5 | 2.9 | -22.0 | 0.34 | -37.7 | -10.4 | | 1.028 |
| M77-2 | Nov 7, 2003 | -11.4 | -83.7 | 164 | -5.9 | 1.7 | -39.6 | 0.23 | | | | |
| M40-4 | July 20, 2005 | -11.1 | -84.4 | 157 | -3.9 | 1.6 | -29.1 | 0.74 | -39.8 | -9.8 | -6.7 | 1.031 |
| M23-2 | July 8, 2004 | -12.0 | -84.7 | 79 | 15.7 | 4.3 | -31.8 | 3.40 | -58.3 | 8.3 | | 1.071 |
| M16-2 | July 21, 2005 | -12.8 | -90.7 | 17 | -12.6 | 0.7 | -26.5 | 0.00 | -59.3 | -16.3 | 1.7 | 1.046 |
| M16-2 | July 8, 2004 | | | 22 | -21.3 | 0.9 | -29.1 | 4.94 | | -21.5 | | |
| M39-4 | July 20, 2005 | -11.9 | -84.2 | 124 | 1.8 | 1.9 | -27.8 | 5.37 | | -5.3 | 7.1 | |
| M39-4 | Nov 7, 2003 | -11.5 | -84.2 | 143 | 2.5 | 5.0 | -18.0 | 2.37 | -57.1 | | | |
| M39-7 | July 20, 2005 | -11.7 | -84.1 | 110 | -1.6 | 2.6 | -27.3 | 1.16 | -31.3 | -8.5 | 7.3 | 1.023 |
| M37-2 | July 20, 2005 | -11.9 | -83.4 | 107 | -0.6 | 4.6 | -26.2 | 5.08 | -61.2 | -7.7 | 5.5 | 1.057 |
| M37-2 | July 8, 2004 | -12.0 | -85.4 | 94 | -4.4 | 7.7 | -18.2 | 2.79 | -61.3 | -11.3 | | 1.053 |
| M37-2 | Nov 6, 2003 | -11.8 | -87.5 | 99 | -2.4 | 8.0 | -31.8 | 1.02 | | | | |
| M4-1 | July 20, 2005 | -12.2 | -89.6 | 132 | 18.3 | 8.5 | -26.9 | 5.79 | -71.2 | 7.2 | 6.0 | 1.084 |
| M4-1 | July 7, 2004 | -12.3 | -85.7 | 129 | 14.9 | 7.0 | -23.3 | 9.88 | -71.4 | 6.5 | | 1.084 |
| M4-1 | Nov 6, 2003 | -12.3 | -89.5 | 127 | 14.4 | 5.0 | -34.9 | 3.42 | | | | |
| M28 | July 22, 2005 | -11.7 | -82.3 | 198 | 22.0 | 13.1 | -25.7 | 6.04 | -53.4 | 11.1 | -9.6 | 1.068 |
| M28 | Nov 6, 2003 | -11.5 | -83.2 | 264 | 22.8 | 23.8 | -37.3 | 0.94 | | | | |
| M90 | Nov 14, 2003 | -11.9 | -84.8 | 86 | 7.0 | 2.8 | -33.4 | 3.09 | | | | |
| M83-2 | Nov 14, 2003 | -10.7 | -80.6 | 136 | -24.2 | 4.0 | -31.7 | 0.22 | | | | |
| Average | | -11.7 | -84.2 | 121 | -1.0 | 4.8 | -29.2 | 2.47 | -51.5 | -7.1 | 1.4 | 1.048 |
| M93 | July 21, 2005 | -11.9 | -85.7 | 60 | -3.3 | 1.6 | -27.0 | 1.81 | -62.3 | -10.1 | 4.5 | 1.056 |
| M93 | July 9, 2004 | -12.1 | -83.8 | 71 | -5.1 | 2.0 | -33.9 | 3.23 | -54.4 | -11.3 | | 1.046 |
| M77-1 | July 21, 2005 | -11.9 | -85.1 | 87 | 10.0 | 3.5 | -26.8 | 3.65 | -71.6 | 2.6 | 7.7 | 1.080 |
| M77-1 | July 9, 2004 | -11.9 | -83.2 | 105 | 9.5 | 5.2 | -30.5 | 5.16 | -59.8 | 2.6 | | 1.066 |
| M77-1 | Nov 7, 2003 | -11.7 | -87.4 | 117 | 9.5 | 4.0 | -34.9 | 1.85 | | | | |
| M114-1 | July 21, 2005 | -11.9 | -87.6 | 80 | -6.6 | 1.2 | -28.9 | 0.95 | -58.2 | -12.0 | 5.7 | 1.049 |
| M114-1 | July 9, 2004 | -12.1 | -83.2 | 71 | -11.2 | 1.3 | -41.9 | 1.04 | -55.9 | -16.6 | | 1.042 |
| M16-1 | July 21, 2005 | -11.6 | -86.9 | 149 | -1.9 | 3.0 | -27.3 | 0.49 | -21.5 | -7.9 | 9.2 | 1.014 |
| M16-1 | July 8, 2004 | -11.7 | -80.1 | 181 | -8.0 | 3.8 | -23.9 | 0.94 | -36.6 | -10.1 | | 1.028 |
| M40-1 | July 20, 2005 | -11.7 | -84.4 | 139 | -4.0 | 2.1 | -27.7 | 0.19 | | -10.2 | 6.4 | |
| M39-1 | July 20, 2005 | -11.9 | -84.4 | 118 | 3.4 | 2.9 | -25.9 | 4.14 | | -4.9 | -9.4 | |
| M39-1 | Nov 7, 2003 | -11.7 | -84.6 | 155 | 3.4 | 3.7 | -36.3 | 3.07 | -52.2 | | | |
| M23-1 | July 20, 2005 | -12.0 | -84.7 | 68 | 13.5 | 2.8 | -27.1 | 6.20 | -67.3 | 4.9 | 7.6 | 1.077 |
| M23-1 | July 8, 2004 | -12.0 | -81.7 | 86 | 11.8 | 4.0 | -37.8 | 5.65 | -55.9 | 5.3 | | 1.065 |
| M107-1 | July 22, 2005 | -11.6 | -83.6 | 26 | -10.8 | 1.0 | -27.3 | 0.01 | 3.6 | -17.4 | -9.0 | 0.979 |
| M37-1 | July 20, 2005 | -12.1 | -86.4 | 28 | -8.6 | 0.6 | -27.6 | 0.32 | -63.3 | -15.1 | 3.5 | 1.051 |
| M37-1 | July 8, 2004 | -12.2 | -84.6 | 30 | -17.5 | 0.8 | -42.0 | 0.16 | -63.7 | -18.9 | | 1.048 |
| M37-1 | Nov 6, 2003 | -12.1 | -89.5 | 34 | -8.5 | 0.8 | -54.2 | 0.19 | | | | |
| M83-1 | Nov 14, 2003 | -11.6 | -84.0 | 170 | 9.3 | 12.5 | -32.1 | 4.72 | | | | |
| M116-1 | Nov 7, 2003 | -10.8 | -80.0 | 59 | 2.1 | 3.3 | -33.0 | 1.41 | | | | |
| Average | | -11.8 | -84.5 | 92 | -0.6 | 3.0 | -32.3 | 2.26 | -51.4 | -7.9 | 2.9 | 1.046 |